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Phosphorus Recovery by Struvite Crystallization from Livestock Wastewater and Reuse as Fertilizer: A Review

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Additional information is available at the end of the chapter

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Abstract

In China, the intensive livestock farming produces massive livestock wastewater with high concentration of phosphorus. Discharge of these compounds to surface water not only causes water eutrophication but also wastes phosphorus resources for plant growth. Therefore, it's necessary combining the removal of phosphorus from livestock wastewater with its recovery and reuse as fertilizer. As a valuable slow-release mineral fertilizer, struvite crystallization has become a focus in phosphorus recovery. In this chapter, struvite crystallization mechanism, reaction factors, crystallizers, and the applications of struvite as fertilizer are discussed. Two steps of nucleation and crystal growth for struvite crystallization from generation to growth are introduced. The reaction factors, including molar ratio of magnesium and phosphate, solution pH, coexisting substances, and seeding assist, of struvite crystallization are summarized. Several innovate types of crystallizer, which relate to the shape and size of harvest struvite to realize the phosphorus recycling, are demonstrated. Due to the influence of toxic or harmful impurities in struvite on its reuse as fertilizer, the environmental risk evaluation of struvite application is introduced. In conclusion, struvite crystallization is a promising tool for recovering phosphorus from livestock wastewater.

Keywords: phosphorus, struvite, livestock wastewater, fertilizer, review

1. Introduction

Phosphorus is a key factor causing water eutrophication, on the other hand, it is also a non-recyclable, nonrenewable, and quite valuable resource. According to the Mineral Commodity Summaries 2015 [1] from the United States Geological Survey (USGS), the reserve of phosphate

rock in China is 3.7 billion tons in 2014, which is in second place in the world. However, with a total of 43–48% of the world's phosphate rock production over the last 3 years [2], the phosphate rock might run out in less than 40 years. So phosphate rock has been one of the 20 minerals that could not meet the demand of the national economy development after 2010 as reported by the Ministry of Land and Resources in China.

On the other side, the intensive livestock farming is a pillar industry in agricultural economy and an important way to increase rural incomes in China [3]. However, it usually produces large amount of livestock wastewater containing high concentration of phosphorus [4]. If this wastewater was not treated reasonably, it would not only lead to the pollution of water eutrophication, but also waste nonrenewable resources and would become one of the major contributors to phosphorus loss [5]. According to the first national sources of pollution survey [6] in China in 2008, nonpoint source pollution in agriculture is a major cause of eutrophication. It accounts for 34.24% of the total phosphorus emission amount, including the livestock and poultry industry. Therefore, it is quite valuable to combine nutrient recycling with environmental pollution control to recover losing phosphorus from livestock wastewater [7].

Numerous phosphorus recovery technologies have been developed, such as biological phosphorus removal, chemical precipitation, electrolysis, adsorption, and crystallization. Biological phosphorus removal utilizes polyphosphate-accumulating organisms to capture phosphorus in their cells. However, this method is limited by the lack of carbon sources and the difficulty of culturing pure bacteria [8]. Chemical precipitation process may consume expensive chemicals and produce large amounts of chemical sludge [9]. Electrolysis is restricted by the small capacity of handling wastewater and the frequent renewal of electrodes [10, 11]. Recovering phosphorus from wastewater using chemical adsorbents is expensive, so cheaper and more efficient adsorbents are necessary for research [12].

Recovering phosphorus by crystallization, by contrast, is a more economical and efficient way. As long as the crystallization conditions are suitable, the struvite crystal would be generated just by adding magnesium (Mg^{2+}) in the raw wastewater which has high concentrations of $H_nPO_4^{n-3}$ and NH_4^+-N . This technology can remove nitrogen at the same time and its production can be used as fertilizer. So it had been studied in many kinds of wastewater, such as multiple wastewater [13], industrial wastewater [14, 15], municipal landfill leachate [16], biogas slurry [17], and effluent of sewage sludge [18], and livestock wastewater is no exception.

2. Struvite characteristic

Magnesium ammonium phosphate, also known as struvite, is a white crystal generated in neutral or mild alkali condition, for which the chemical formula is $Mg(NH_4)PO_4 \cdot 6H_2O$. Struvite consists of one molecule of magnesium (Mg^{2+}), one molecule of ammonium (NH_4^+),

one molecule of phosphate (PO_4^{3-}), and six molecules of water (H_2O), whose relative molar mass is 245.43 g/mol. It is only slightly soluble in water but soluble in acid solution [19]. Struvite is a light crystal with low relative density of 1.65–1.7. It is not easy to be rush off by rainfall [20]. Pure struvite belongs to orthorhombic crystals which consists of regular PO_4^{3-} octahedron, distorted $\text{Mg}(\text{H}_2\text{O})_6^{2+}$ octahedron, and groups of NH_4^+ connected by hydrogen bonding [21], but shows rod-like structure [22] or irregular structure [23] sometimes (**Figure 1**). And struvite of rod-like structure is of low purity, because of the coprecipitation with foreign ions.

Actually, struvite had been widely studied as early as 1937 for the congestion in the pipes of the sludge anaerobic digester [24]. The general struvite forming reaction equation is shown below:



When Mg^{2+} , NH_4^+ , and $\text{H}_n\text{PO}_4^{n-3}$ ($n = 0, 1$, or 2) exist in the solution and the product of their concentrations are bigger than the solubility product constant (K_{sp}) of struvite, the crystal would be generated spontaneously. And the calculation formula of struvite's K_{sp} is shown below:

$$K_{\text{sp}} = [\text{Mg}^{2+}] \cdot [\text{NH}_4^+] \cdot [\text{PO}_4^{3-}] \quad (4)$$

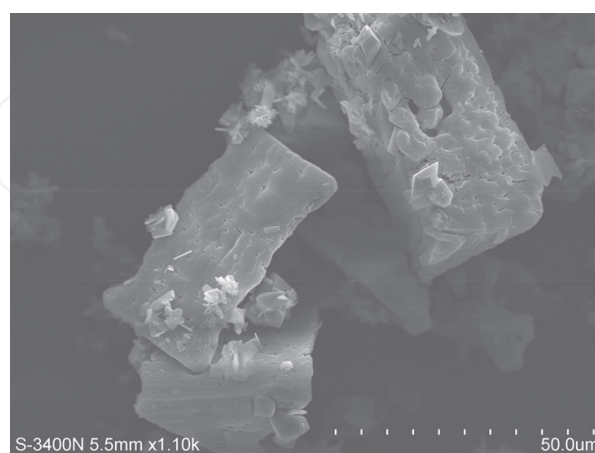


Figure 1. The SEM figure of magnesium ammonium phosphate crystal.

where $[Mg^{2+}]$, $[NH_4^+]$, and $[PO_4^{3-}]$ are concentrations of Mg^{2+} , NH_4^+ , and PO_4^{3-} in the solution, respectively. As the molar ratio of Mg^{2+} , NH_4^+ , and PO_4^{3-} is 1:1:1 in struvite, so C^* is used to present the same concentration of these three ions, which means $C^* = [Mg^{2+}] = [NH_4^+] = [PO_4^{3-}]$. So the calculation formula of struvite's K_{sp} also can be shown as:

$$K_{sp} = (C^*)^3 \tag{5}$$

Snoeyink et al. [25] got the K_{sp} of struvite is $10^{-12.6}$ as early as 1980. Ohlinger et al. [26] corrected it to $10^{-13.26}$ in 1999. And then Bhuiyan et al. [27] corrected it again to $10^{-13.36}$ in 2007, which is widely used now. However, K_{sp} of struvite is hard to get in the real wastewater for the negative impact of the soluble coexisting ions. Therefore, in the estimation of the saturability of the real wastewater, ionic activity coefficient (K_{so}) is more widely useful than K_{sp} . Considering the impact of ionic strength (I) and the ionic activity (A_i) in estimating the K_{so} , the value of K_{so} is bigger than K_{sp} . And the calculation formula of struvite's K_{so} is shown below:

$$K_{so} = \alpha_{Mg^{2+}} \times \alpha_{NH_4^+} \times \alpha_{PO_4^{3-}} \tag{6}$$

$$\alpha_i = \gamma_i [C_i] \tag{7}$$

where α_i presents the ionic activity (A_i), γ_i presents the activity coefficient of the ionic strength (I), and $[C_i]$ presents the concentration of the ion. Only when the value of γ_i is 1, K_{sp} is able to represent K_{so} . Therefore, it is necessary to eliminate the interruptions of the soluble coexisting ions (like Ca^{2+} , CO_3^{2-} , and SO_4^{2-}) and clear of the ionic activities of Mg^{2+} , NH_4^+ , and PO_4^{3-} in the specific pH condition. **Table 1** shows the ionization equations and pKa value in magnesium ammonium phosphate solution at 25°C, which is helpful to estimate the distribution of these ions and predict the probability to generate struvite under such environment of solution.

| No. | Ionization equation | pK _a |
|-----|--|-----------------|
| 1 | $NH_4^+ \rightleftharpoons NH_3(aq) + H^+$ | 9.26 |
| 2 | $H_3PO_4 \rightleftharpoons H_2PO_4^- + H^+$ | 2.12 |
| 3 | $H_2PO_4^- \rightleftharpoons HPO_4^{2-} + H^+$ | 7.20 |
| 4 | $HPO_4^{2-} \rightleftharpoons PO_4^{3-} + H^+$ | 12.36 |
| 5 | $MgNH_4PO_4 \cdot 6H_2O \rightleftharpoons Mg^{2+} + NH_4^+ + PO_4^{3-} + 6H_2O$ | 12.70 |
| 6 | $MgOH^+ \rightleftharpoons Mg^{2+} + OH^-$ | 2.56 |
| 7 | $MgH_2PO_4^+ \rightleftharpoons H_2PO_4^- + Mg^{2+}$ | 0.45 |
| 8 | $MgHPO_4 \rightleftharpoons HPO_4^{2-} + Mg^{2+}$ | 2.91 |
| 9 | $MgPO_4^- \rightleftharpoons PO_4^{3-} + Mg^{2+}$ | 4.80 |

Table 1. The ionization equations and pKa value in magnesium ammonium phosphate solution at 25° [30, 31].

3. Mechanism of struvite crystallization

Nucleation and crystal growth are two classical steps in the process of struvite crystallization from generation to development. As shown in **Figure 2**, nucleation is the first step of the struvite crystallization. When Mg^{2+} , NH_4^+ , and PO_4^{3-} meet under the proper pH value, nucleation occurs. And the nucleation time is the time required to form a saturated solution to the beginning of the nucleation. It is mainly influenced by the pH of solution, mixing energy, coexisting ions, and saturation index (SI). The ion activity affected by the value of pH significantly leads to differentiation of combine speed of free ions [28]. Weak ion activity means slow combine speed and longer nucleation time indirectly. When the rate of struvite nucleation and growth is greater than or equal to the rate of mixing magnesium to the solution, there needs additional mixing energy. Kim et al. [15] emphasized that mixing energy could influence the quantity and size of struvite strongly. However, the greatest impact of struvite nucleation is saturation index (SI) of solution which decides the development of crystal to homogeneous or heterogeneous directly [26]. SI is used to describe the saturation state of the reaction system of struvite. And the SI calculation of struvite is shown as follows:

$$SI = \log \frac{IAP}{K_{sp}} \quad (8)$$

where IAP and K_{sp} represent ionic activity product and the thermodynamic solubility product of struvite, respectively [29]. The homogeneous crystallization that we want happens on metastable region in the solution. In this region, nucleation is not spontaneous, which differentiated between the process of crystallization and precipitation, and avoids the occurrence of undesirable spontaneous nucleation to a great extent [28]. However, metastable state of solution is very difficult to control. Therefore, SI, as the indicator for metastable state, is very important. Bonurophoulos et al. [30] found that the threshold between homogeneous and heterogeneous precipitation is the condition where $SI \approx 2.0$ and the nucleation rate of 1 nucleus/($\text{cm}^3 \cdot \text{s}$). When the SI is less than 1.716, the struvite crystals are in heterogeneous precipitation and vice versa. Bhuiyan et al. [31] and Mehta et al. [32] also got the threshold at $SI = 1.83$ and $SI = 1.7$ at the special nucleation rate, respectively. In addition, Durrant et al. [33] emphasized the great influence of SI on the shape of struvite as early as 1999. And it also has a SI threshold between rhombic structure and rod-like structure of struvite.

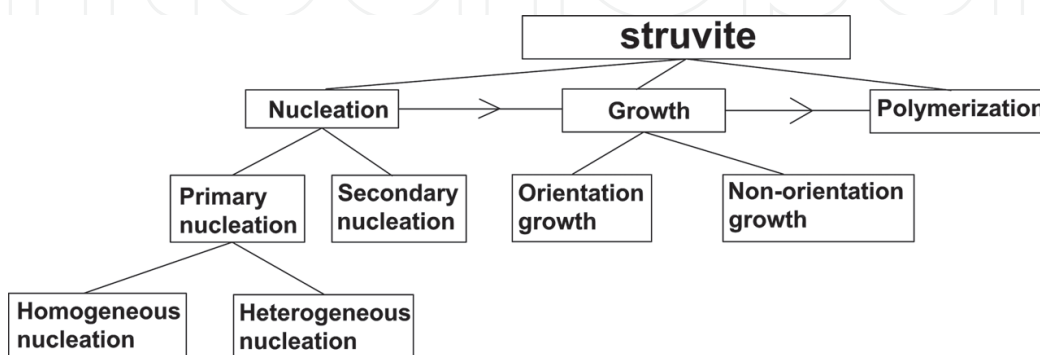


Figure 2. The crystal nucleation, growth and aggregation mechanism of magnesium ammonium phosphate.

After the crystal nucleus generates, the ions in the solution used to form the crystal begin to deposit on the crystal nucleus and the nucleus grow to the settling particles. During that time, there are two trends for the development of particles. One is orientation growth, which means the ions sequence in the crystal is arranged according to a certain lattice. The other one is nonorientation growth, which means these ions are too late to arrange in order. It is the fast growth rate that causes disorder. And two types of the crystal growth mechanisms lead to different trends. One is the integration mechanism, and the other is the mass transfer mechanism. The former is the integration of solute molecules into the surface; the latter is the transfer (by diffusion or convection) of solute molecules from the bulk solution to the crystal surface. When the effect of mass transfer is greater than the effect of integration, the crystal growth mainly depends on the diffusion effect and the growth of crystal would be orientable. However, if the effect of integration is greater, the integration on the surface of solute decides the crystal growth. And the relative sizes of the nucleation rate, aggregation rate, and directional array rate also decide the trend of crystal growth, which can be changed by precipitation conditions [31]. Abe et al. [34] showed that the growth rate of struvite was very slow. In the high concentration of phosphate (greater than 200 mg/L), the daily growth rate of struvite was 0.173 mm. In the low concentration of phosphate (30–100 mg/L), the daily growth rate of struvite was 0.061 mm. Therefore, increasing crystal growth rate and crystal size of struvite is not only beneficial to further removing phosphorus from livestock wastewater, but also to recycling phosphorus with a bigger size struvite. There is a metastable zone in industrial crystallization to make the crystal bigger and more even. The metastable zone is defined as a region bounded by the solubility curve in which the solution is supersaturated but the spontaneous nucleation cannot occur in such a short time [35]. In the metastable zone, the solute condenses on the nucleus as constantly as possible. As we known, the process of struvite constant growth is also the further recovery of phosphorus from livestock wastewater. So it is meaningful to study the metastable zone of struvite for the industrialized application.

4. The factors influencing struvite crystallization

4.1. Molar ratio of P and Mg

Generally speaking, livestock wastewater is rich in ammonium and phosphorus. So it is needed to add extra magnesium to form struvite. Therefore, the addition amount of magnesium affects the solubility product constant (K_{sp}) directly, which further affects the quantity of struvite crystal and the recovery rate of phosphorus in livestock wastewater. So the molar ratio of phosphate and magnesium is the key factor to control the yield of phosphorus recovery. The molar ratio of phosphate and magnesium is 1:1 in theory. However, the real molar of the added magnesium is larger than the total amount of phosphorus in the real livestock wastewater. As shown in **Table 2**, for a higher phosphorus removal rate, the molar ratio of phosphate and magnesium is about 1:1–1.2 from livestock wastewater, 1:1.4 from synthetic livestock wastewater, and 1:1–1.4 from anaerobic digesters of livestock wastewater. It is mainly based on the effect of coexisting ions in the livestock wastewater. The coexisting ions, such as OH^- and CO_3^{2-} , are apt to coprecipitate with Mg^{2+} , which prevent the Mg^{2+} from touching

| Samples | Initial concentration of phosphate (mg/L) | Molar ratio of N, P, and Mg | pH | Reaction time | Removal rate of phosphate (%) | References |
|--|---|-----------------------------|----------|---------------|-------------------------------|------------|
| Animal manure wastewater | 145 | 16.4:1:1.05 | 8 | 30 min | 67 | [40] |
| Animal manure wastewater | 189.9 | 1:1:0.8–1 | 8.35 | 4 h | 96 | [41] |
| Animal manure wastewater | 60.01 | 63.5:1:1 | 8.09 | 4 h | 92.82 | [42] |
| Animal manure wastewater | 128 ± 13 | 1:1:1.2 | 9 | 1 h | 98 | [43] |
| Synthetic animal manure wastewater | 80 | 8:1:1.4 | 9.5–10.5 | 2 h | 97 | [44] |
| Synthetic animal manure wastewater | 130.2 | 1:1:5 | 7.9 | – | 92 | [45] |
| Anaerobic digesters of manure wastewater | 51.1 | 30.7:1:1.4 | 8.0–10.0 | 1 h | 74–95 | [46] |
| Anaerobic digesters of manure wastewater | 55.4 | 9.6:1:1.2 | 9.0 | 20 min | 85 | [47] |
| Anaerobic digesters of manure wastewater | 64.2 | 1:1.2:1.2 | 9.0 | 15 h | 97.2 | [48] |

Table 2. The summary of parameters on magnesium ammonium phosphate crystallization.

with the NH_4^+ and PO_4^{3-} , so more magnesium is needed. Marit et al. [36] indicated that there would be many other kinds of magnesium phosphate precipitates except for struvite at different value of pH, such as $\text{Mg}_3(\text{PO}_4)_2$, MgHPO_4 , and $\text{Mg}(\text{H}_2\text{PO}_4)_2$. So it does not mean that more magnesium means more struvite. Excessive amounts of magnesium would increase the pH value of the solution as well as the degree of saturation of the magnesium salts, resulting in the formation of other kinds of magnesium phosphate precipitates mentioned.

4.2. The wastewater pH value

The value of pH of livestock wastewater is an important parameter to the formation of struvite. It affects the quality and the purity of struvite at the same time. As shown in **Table 2**, the best value of pH to form struvite is between 8 and 10, while 8.0–9.0 is the best for livestock wastewater. However, Hao et al. [37] indicated that struvite could get the highest purity at pH = 7.0, and the purity seemed to have fallen with the increasing of the pH value of wastewater. When the value of pH is higher than 10, the formed precipitate mainly consists of $\text{Mg}_3(\text{PO}_4)_2$ ($K_{\text{sp}} = 9.8 \times 10^{-25}$). Song et al. [38] also found that the precipitate of $\text{Mg}(\text{OH})_2$ would form at the pH of 11. It does not mean that it is better to form struvite at a lower value of pH. However, considering the phosphorous recovery, as long as the productions of phosphorus salts are harmless and nontoxic, the aim of recovering phosphorus from livestock wastewater is reached. Anyway, the pH value of livestock wastewater is generally between 7.5 and 8.5, which is more convenient to recover phosphorus without the need for adjusting the pH value. It is helpful to simplify the technology and reduce the cost of livestock wastewater treatment.

4.3. The coexisting ions

There are many kinds of coexisting ions interfering with the crystallization of struvite in livestock wastewater, such as calcium ion (Ca^{2+}), carbonate ion (CO_3^{2-}), suspended solids (SS), and heavy metal ions (HMI). Moerman et al. [39] found that Ca^{2+} could enhance the phosphorus removal with forming the precipitate of $\text{Ca}_3(\text{PO}_4)_2$. However, it reduces the size of struvite. Meanwhile, lots of $\text{Ca}_3(\text{PO}_4)_2$ powder flows out with effluent easily, declining the effluent water quality. Le Corre et al. [40] also declared that Ca^{2+} would compete with Mg^{2+} and form the precipitates of $\text{Ca}_3(\text{PO}_4)_2$ ($K_{\text{sp}} = 2.1 \times 10^{-33}$) and CaHPO_4 ($K_{\text{sp}} = 1.8 \times 10^{-7}$) at the pH value of 9. By performing batch experiments, Zhang et al. [41] found that the degree of the supersaturation would decrease with the increase of the concentration of CO_3^{2-} . The CO_3^{2-} , easily combining with Mg^{2+} , increases the ion saturation in the solution and decreases the concentration of Mg^{2+} forming struvite. Suzuki et al. [42] showed that negatively charged SS adsorbed NH_4^+ and Mg^{2+} easily in the alkaline environment, which retarded the struvite crystalline rate. And Muryanto et al. [43] studied on the influence of copper ions (Cu^{2+}) and zinc ions (Zn^{2+}) in struvite crystallization and showed that the existence of Cu^{2+} and Zn^{2+} would delay the nucleation rate and the growth rate of struvite. Although they had little impact on the crystal shape, the crystal would have some cracks on the surface.

All in all, in the process of recovering phosphorus from livestock wastewater, some pretreatments are necessary to implement for removing these coexisting ions before forming the struvite. Laridi et al. [44] tried to reduce the negative impacts of organics and SS by adding ferric chloride and flocculants into the livestock wastewater, and it worked with a higher phosphorus removal rate at the same time. Suzuki et al. [42] tried to separate the struvite from suspended solids containing heavy metals by the differences of their settlement characteristics. It improved the purity of struvite and reduced the negative impact of SS and heavy metal ions.

4.4. Seed crystal

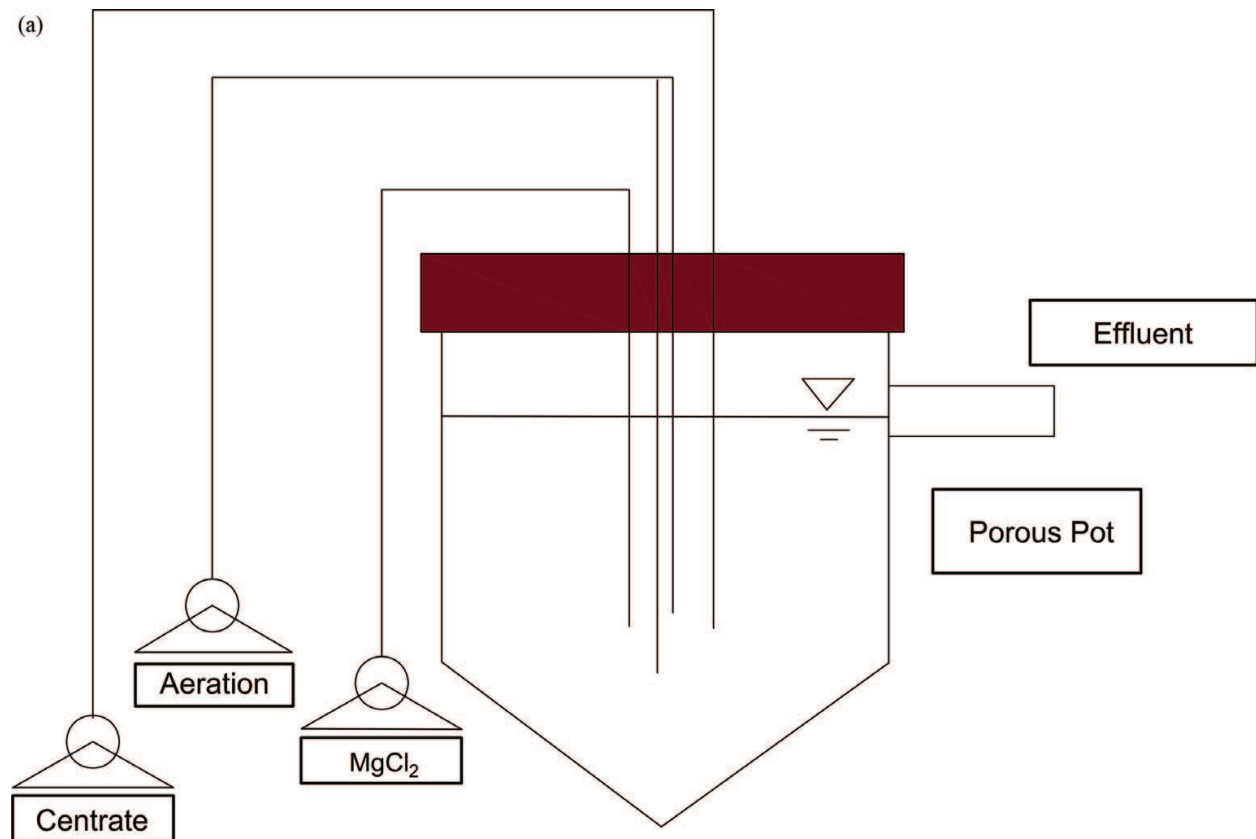
Seed crystals have positive influence in the struvite growth. Adding seed crystals can reduce the saturation of struvite crystallization in need, shorten the nucleation time, and increase the rate of crystal growth. What is more, struvite crystallizes on the surface of seed crystals, which enhances the separation of crystals and water, prevents the tiny crystals from flowing out with the effluent, and improves the phosphorus removal efficiency. Ariyanto et al. [45] showed that the smaller the added crystal nucleus is, the faster is the rate of crystal growth. Kim et al. [18] emphasized that the excessive amount of seed crystals added could not improve the phosphorus removal efficiency, and the pH value of wastewater also influenced the efficiency at the same dosage of seed crystals. The phosphorus removal efficiency is more significant at the pH value of 9. So only adding proper amount of seed crystals with a proper average size can the phosphorus removal efficiency be higher.

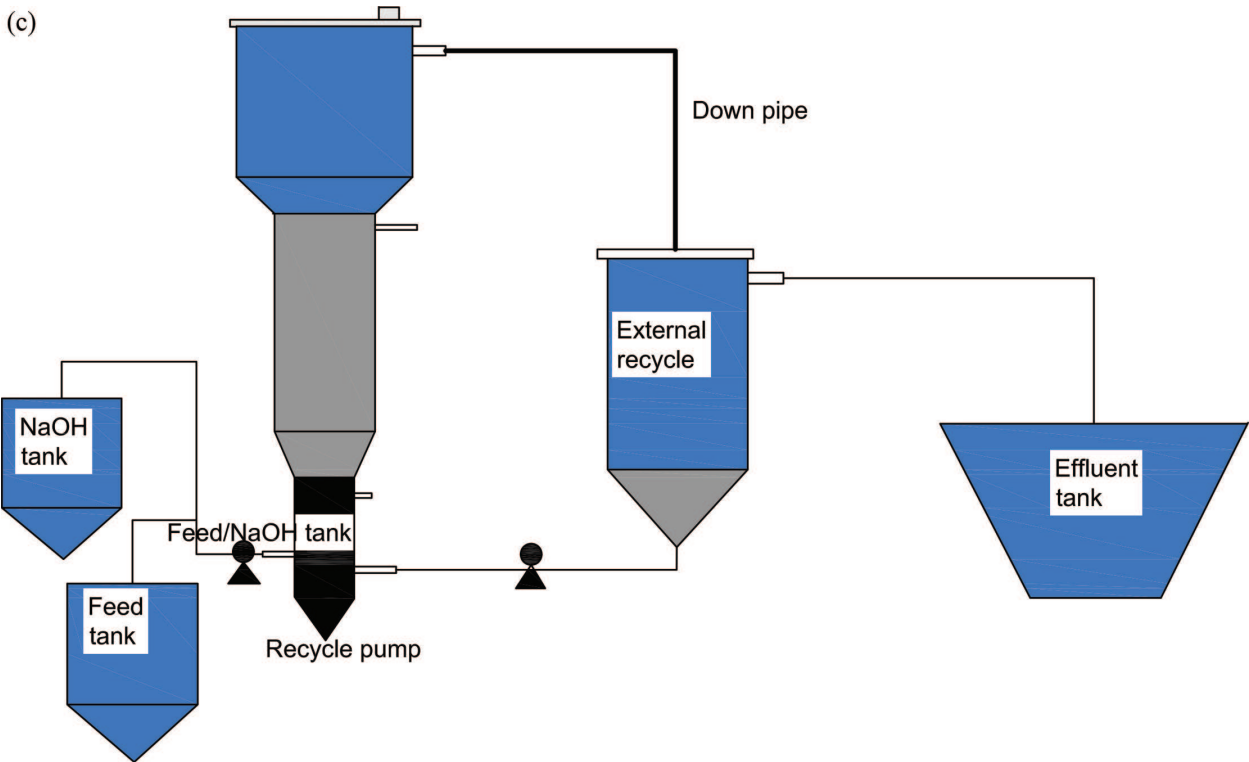
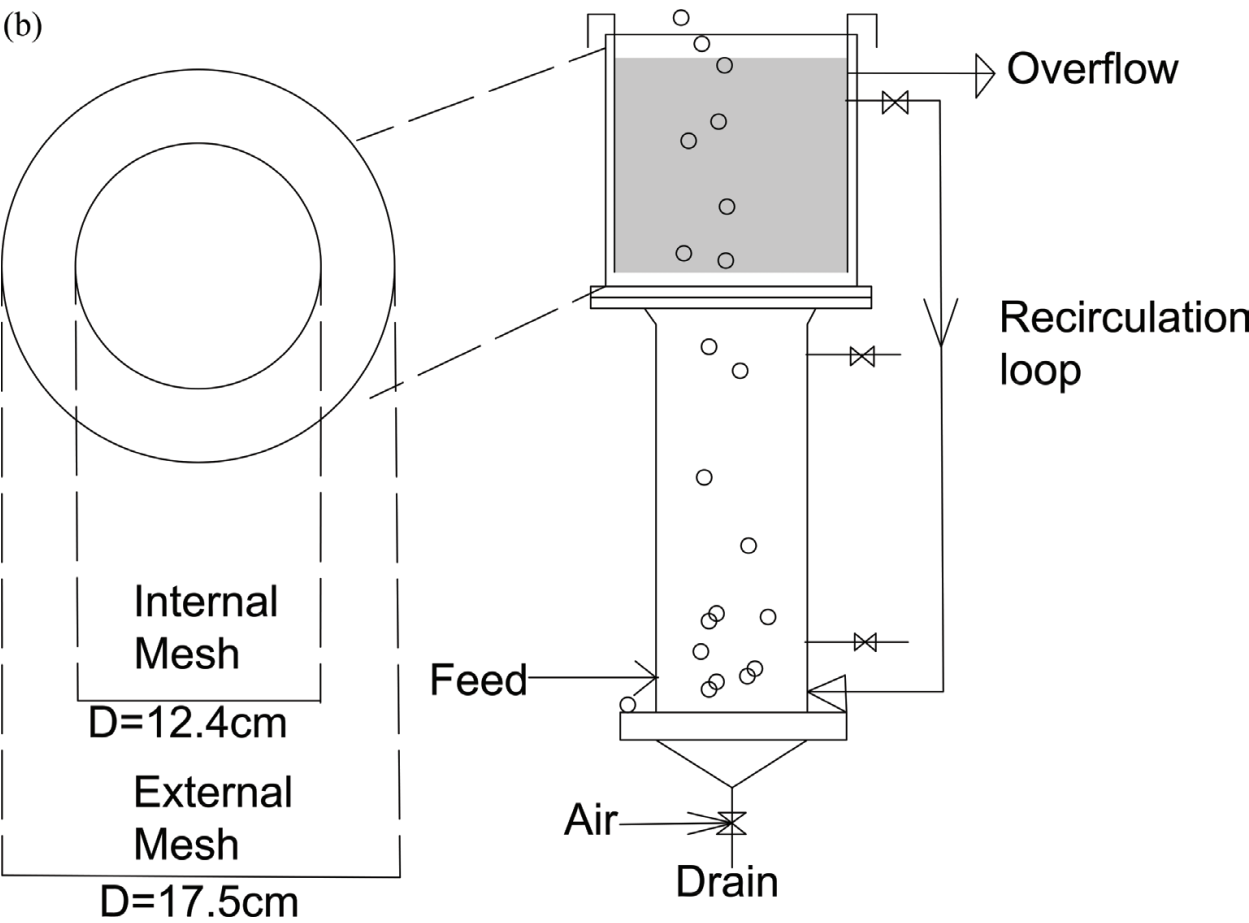
5. Crystallizer of struvite

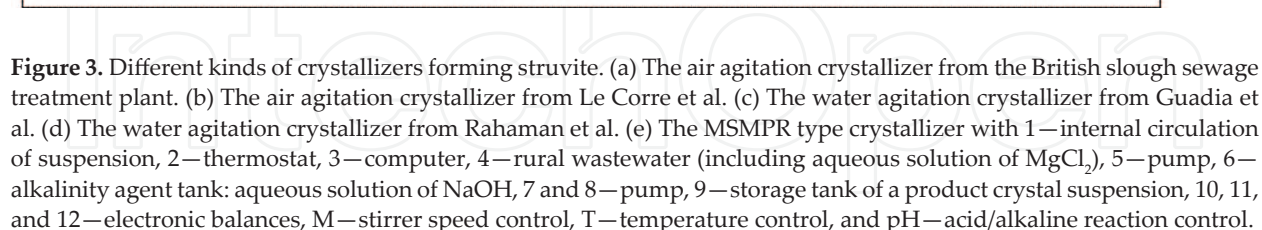
It is important to realize the phosphorus recycling in crystallizer, as the struvite crystallization equipment. The pros and cons of its design decide the shape and size of the struvite and

the phosphorus removal efficiency from livestock wastewater. A series of struvite crystallizers had been developed and put into production successfully abroad previously, which had obtained environmental and economic benefits simultaneously. According to the mode of agitation, these crystallizers can be divided into air agitation type, water agitation type, and mechanical agitation type.

The air agitation type crystallizer is a kind of crystallizer that is studied widely. The special aerate system can not only mix the solution more efficiently, improving the collision chance of crystal forming ions, but also vent the gas carbon dioxide and insoluble ammonium from the solution, increasing the pH value of the solution, the ammonium removal efficiency, and the effluent water quality at the same time. The crystallizer used by the British slough sewage treatment plant (**Figure 3a**) reached the soluble phosphorus removal rate of 94% and the total phosphorus removal rate of 87.5% with drumming into air under the inner reaction zone [46]. Le Corre et al. designed two concentric meshes made of stainless steel as a substrate to grow struvite in the crystallizer, which can trap and then accumulate the struvite in the reactor as an adhesive (**Figure 3b**). With the help of crystallizers, the phosphorus removal rate can increase from 81 to 86%. However, because of the limitations of volume and growth time, the struvite crystal cannot grow large enough in the air agitation type crystallizer, which causes the loss of phosphorus recovery. Moreover, some kinds of air agitation type crystallizer have the problem of replacing padding or membranes frequently. What is more, the congestion problem becomes serious once the crystallizer broke down for some reasons, and it is hard to restart.







hopper) at the top, getting the phosphate removal rate of up to 90% and the size of particles up to 3.5 mm. Seed crystals are added into the crystallizer from the seed hopper and allowed the finer crystals to continue to grow up in the upper supersaturated solution.

The mechanical agitation type crystallizer, mixing solution by impellers, is simple in design and easy to operate. However, it causes greater energy consumption and uneven size of crystals distribution. Recently, a new crystallizer called mixed suspension, mixed product removal crystallizer (MSMPR for short) can solve the problem of uneven crystal size distribution (**Figure 3e**). With the mechanical agitation centered and the water agitation assisted, MSMPR can uniform the suspension density and particle size of the crystals, and remove productions evenly by controlling the speed and time of mixing. Hutnik et al. [48] and Kozik et al. [49] both got the phosphate removal rate up to 99% from industrial wastewater and wastewater with low concentration of phosphorus. And both of them confirmed that MSMPR could increase the size of crystals, improve the crystallization rate of the struvite, and enhance the phosphorus recovery rate.

6. Application of struvite as a fertilizer

It is reckoned that 100 m³ wastewater could form 1 kg of struvite. If all the wastewater in the world is treated by struvite crystallization, 63,000 tons of P₂O₅ could be recovered, equaling to 16% of the phosphate rock production of the world [50]. And 171 g struvite can be recovered from livestock wastewater per square meter at most and the purity as high as 95% without washing. Therefore, recovery of struvite returning to the farmland is a developmental trend of struvite crystallization technology. Struvite, as a slightly soluble crystal, for containing the equal molar concentrations of magnesium (Mg), ammonium, and phosphate, has been successfully used on herbages [51], vegetables [14, 52], and grain crops [53] as a fertilizer, especially on the magnesium-fond crops, like sugar beet [54]. Especially, the presence of Mg in struvite makes it more attractive as an alternative to contemporary fertilizers for a few crops, which require magnesium [55]. Ryu et al. [52] found that the struvite source provided the essential crop nutrients of N, P, K, Ca, and Mg for Chinese cabbage as much as other commercial fertilizers. Moreover, it has a lasting positive function to roots and does not burn the seeding or roots due to its slow release characteristics. Besides, compared with other highly soluble fertilizers, struvite is more suitable for use in the vast areas of forest. Since the area of forest is too large to fertilize frequently, the use of struvite can decrease the frequency of fertilization and reduce the loss of nutrients [54]. However, as livestock wastewater is full of impurities, especially the heavy metal ions, the struvite recovered from livestock wastewater still contain more or less heavy metal ions. From livestock-based struvite, toxic substances may diffuse into the aquatic environment or accumulate in soils and have an adverse effect on the human health and environment [56, 57]. Although currently no specific threshold values are available for micropollutants in fertilizers, the introduction of potential hazardous substances into the environment should be avoided. The accumulation of heavy metal ions will be a serious concern for sustainability [58]. Ryu et al. [52] made a security evaluation for

struvite as a fertilizer used in the soil. They affirmed the fertilizer efficiency of struvite and emphasized the negative effect of higher concentrations of copper and cadmium in struvite at the same time. Because copper and cadmium were tested in the cabbage fertilized by the struvite used as fertilizer, the struvite, especially recovered from livestock wastewater, needs to be tested for the amount of toxic or harmful substances, followed by the security evaluation as a fertilizer.

7. Summary and outlook

Struvite crystallization represents a promising tool for recovering phosphorus from livestock wastewater. Based on this study, the conclusions are as follows.

Struvite is a white crystal which is formed in neutral or mild alkali conditions. Nucleation and crystal growth are two steps for struvite crystallization from generation to growth. The molar ratio of magnesium and phosphate, and solution pH are the key factors to control. The coexisting substances, such as calcium, carbonate, suspended solids, and heavy metals, interfere the crystallization of struvite. Seed crystals have positive influence on struvite growth. Adding seed crystals can reduce the saturation of struvite crystallization, shorten the nucleation time, and improve the rate of crystal growth.

Crystallizer, its design decides the shape and size of struvite, is important to realize the phosphorus recycling. According to the agitation mode, it can be divided into air agitation, water agitation, and mechanical agitation.

The recovered struvite can be used on herbages, vegetables, and grain crops as a fertilizer, especially on the magnesium-fond crops, like sugar beet.

However, there are still some problems. Livestock wastewater belongs to the organic wastewater with high concentrations of ammonium, phosphorus, organics, and suspended solids. And the existing forms are complex, such as simple monoester phosphorus, phytate-like phosphorus, and polynucleotide-like phosphoric. So it is necessary to use some physical or chemical measures to transform different kinds of phosphorus to phosphate, as many as possible, before removing phosphorus from livestock wastewater.

The design for struvite crystallizer is still a key to struvite crystallization technology. Although enhancing the phosphorus removal rate has got a big breakthrough by the current crystallizer, the crystals are still too small to recover and block the crystallizer easily. It is the influence of negative zeta potential on the surface of crystals that makes further aggregating hard for small crystals. So finding a way to change the zeta potential on the surface of the crystals, enhancing the aggregation capability, and increasing the size of the crystals is required.

Struvite, a fertilizer with high concentrations of nutrients, might be difficult in application for the influence of other toxic or harmful impurities. Therefore, to reduce the environmental risk at source, it is necessary that estimating the potential effects of struvite on the ecosystem before use.

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